



Background document for Nucella lapillus (Dog whelk)



OSPAR Convention

The Convention for the Protection of the Marine Environment of the North-East Atlantic (the "OSPAR Convention") was opened for signature at the Ministerial Meeting of the former Oslo and Paris Commissions in Paris on 22 September 1992. The Convention entered into force on 25 March 1998. It has been ratified by Belgium, Denmark, Finland, France, Germany, Iceland, Ireland, Luxembourg, Netherlands, Norway, Portugal, Sweden, Switzerland and the United Kingdom and approved by the European Community and Spain.

Convention OSPAR

La Convention pour la protection du milieu marin de l'Atlantique du Nord-Est, dite Convention OSPAR, a été ouverte à la signature à la réunion ministérielle des anciennes Commissions d'Oslo et de Paris, à Paris le 22 septembre 1992. La Convention est entrée en vigueur le 25 mars 1998. La Convention a été ratifiée par l'Allemagne, la Belgique, le Danemark, la Finlande, la France, l'Irlande, l'Islande, le Luxembourg, la Norvège, les Pays-Bas, le Portugal, le Royaume-Uni de Grande Bretagne et d'Irlande du Nord, la Suède et la Suisse et approuvée par la Communauté européenne et l'Espagne.

Acknowledgement

This report has been prepared by the Netherlands as lead country for Nucella lapillus.

Photo cover page: Dog whelk@Jakob Strand

Contents

Backgr	ound document for Nucella lapillus (Dog whelk)	4
Exe	ecutive Summary	4
Réd	capitulatif	4
1.	Background information	5
2.	Original evaluation against the Texel-Faial selection criteria	5
	OSPAR Regions where the species is under threat and/or in decline	5
	Original evaluation against the Texel-Faial criteria for which the species wa	ıs
	included on the OSPAR List	5
	Relevant additional considerations	6
3.	Current status of the species	6
	Distribution of the feature in the OSPAR maritime area	6
	Population (current/trends/future prospects)	7
	Condition (current/trends/future prospects)	7
	Limitations in knowledge	8
4.	Evaluation of threats and impacts	
	Threat and link to human activities	9
5.	Existing management measures	10
6.	Conclusions on overall status	11
7.	What action should be taken at an OSPAR level?	12
	Action/measures that OSPAR could take, subject to OSPAR agreement	
	Brief Summary of the proposed monitoring system	12
Annex	1. Overview of data and information provided by Contracting Parties	13
Sur	mmaries of country-specific information provided	15
Annex	2. Description of the recommended monitoring and assessment strategy	17
Rat	ionale	17
Use	e of existing monitoring programmes	17
Syr	nergies with monitoring of other species or habitats	17
Ass	sessment criteria	18
Tec	chniques/approaches:	19
Sel	ection of monitoring locations	19
Tim	ning and Frequency of monitoring	20
Dat	a collection and reporting	20
Qua	ality assurance	20
Δηηργ	3 References	21

Background document for *Nucella lapillus* (Dog whelk)

Executive Summary

This background document on the Dog whelk – *Nucella lapillus* - has been developed by OSPAR following the inclusion of this species on the OSPAR List of threatened and/or declining species and habitats (OSPAR other agreement 2008-6). The document provides a compilation of the reviews and assessments that have been prepared concerning this species since the agreement to include it in the OSPAR List in 2003. The original evaluation used to justify the inclusion of *Nucella lapillus* in the OSPAR List is followed by an assessment of the most recent information on its status (distribution, population, condition) and key threats prepared during 2008-2009. Chapter 7 provides recommendations for the actions and measures that could be taken to improve the conservation status of the species. On the basis of these recommendations, OSPAR will continue its work to ensure the protection of *Nucella lapillus*, where necessary in cooperation with other organisations. This document may be updated to reflect further developments.

Récapitulatif

Le présent document de fond sur la *pourpre petite pierre* a été élaboré par OSPAR à la suite de l'inclusion de cette espèce dans la liste OSPAR des espèces et habitats menacés et/ou en déclin (Accord OSPAR 2008-6). Ce document comporte une compilation des revues et des évaluations concernant cette espèce qui ont été préparées depuis qu'il a été convenu de l'inclure dans la Liste OSPAR en 2003. L'évaluation d'origine permettant de justifier l'inclusion de la *pourpre petite pierre* dans la Liste OSPAR est suivie d'une évaluation des informations les plus récentes sur son statut (distribution, population, condition) et des menaces clés, préparée en 2008-2009. Le chapitre 7 recommande des actions et mesures à prendre éventuellement afin d'améliorer l'état de conservation de l'espèce. OSPAR poursuivra ses travaux, en se fondant sur ces recommandations, afin de s'assurer de la protection de la *pourpre petite pierre* le cas échéant en coopération avec d'autres organisations. Le présent document pourra être actualisé pour tenir compte de nouvelles avancées.

1. Background information



N.lapillus is a gastropod mollusc that is found on wave exposed to sheltered rocky shores. It is widely distributed on both sides of the North Atlantic where there is suitable habitat. In the OSPAR Maritime Area, its distribution extends from Iceland in the north, to Portugal in the south and includes Irish Sea and North Sea coasts.

2. Original evaluation against the Texel-Faial selection criteria

OSPAR Regions and Dinter biogeographic zones where the species occurs

OSPAR Regions: All

Dinter Biogeographic zones: Warm-temperate pelagic waters, Lusitanean (Cold/Warm)

Lusitanean-boreal, Boreal-lusitanean, Boreal, Norwegian Coast

(Skagerrak)

OSPAR Regions where the species is under threat and/or in decline

Regions specified for decline and/or threat: II, III, IV

Original evaluation against the Texel-Faial criteria for which the species was included on the OSPAR List

N.lapillus was nominated for inclusion on the OSPAR list based on the criteria for decline and sensitivity, with information also provided on threat.

Decline: *N.lapillus* populations are known to have declined in certain locations throughout their range in the OSPAR Maritime Area. They used to be very common on the coast of Belgium but disappeared during the end of the 1970s and early 1980s (Kerckhof, 1988). In the UK, local declines have been reported by Bryan *et al.*, (1986) in south-west England. The decline has been linked to contamination effects of tributyltin (TBT) compounds used in antifouling paints, which cause imposex in *N.lapillus* (see section on threats). Evans *et al.*, (1996) concluded, for example, that the extinction of several populations in the UK and the east coast of the North Sea were due to TBT contamination.

Sensitivity: An assessment of the sensitivity of *N.lapillus*, based on a literature review by the Marine Life Information Network for Britain & Ireland (MarLIN), lists this species as being highly sensitive to synthetic compound contamination, changes in nutrient levels, and substratum loss (Tyler-Walters,

2002). The most extensively studied sensitivity is in relation to TBT, which is known to cause an irreversible condition known as 'imposex' where females develop male characteristics. The effects can be seen at very low concentrations. Imposex in *N.lapillus* is fully developed at ambient TBT concentrations of 1-2 ng/l and at 3 ng/l or more females are fully sterilised (Gibbs & Bryan, 1996). The percentage of females in a locality falls with increasing degree of imposex, which puts additional pressure on the population (Bryan *et al.*, 1986). Sensitivity to changes in nutrient levels have been described by Gibbs *et al.* (1999) who reported a massive kill of *N.lapillus* in Bude Bay, north Cornwall, and suggested that the mass mortalities may have been caused by eutrophication and summer algal blooms linked to a new sewage outfall in the area. *N.lapillus* has also been shown to be severely affected by toxic algal blooms. These have been reported from south-west Ireland following a bloom of Gyrodinium aureolum in 1979 (Cross & Southgate, 1980), a bloom of Chrysochromulina polylepis in the Kattegat, Skagerrak and Norwegian coast of the North Sea in 1988 (Underdal *et al.*, 1989), and up to 98-99% mortality of *N.lapillus* exposed to a toxic bloom of Chrysochromulina polylepis in Gullmar Fjord, west Sweden in June 1988 (Robertson, 1991).

Threat: Imposex in *N.lapillus*, which has been linked to exposure to TBT from antifouling paints, is one of the most widely reported threats to *N.lapillus* in the OSPAR Maritime Area. It was first recognised in *N.lapillus* by Blaber (1970) in *N.lapillus* collected from the south coast of England. Significant changes were also noted between its prevalence in the late 1960s and 1985, with the prevalence of imposex rising from 5% and less than 0.1% at two sites studied, to 67% and 48% respectively. The effects of TBT have since been observed in *N.lapillus* from the coastal areas of all countries bordering the North Sea, the Atlantic coast of Spain and Portugal, as well as in the more remote northern shores around Iceland (OSPAR 2000; Harding et al., 1999; Svavarsson and Skarphéoinsdóttir, 1995; Skarphédinsdóttir et al., 1996). It has been accepted that imposex is induced almost typically by TBT used in antifouling paints, based on results of laboratory experiments using *N.lapillus* (Bryan et al., 1986). It has been suggested that the high levels of imposex in *N.lapillus* around marine European shipping and fishing ports are unlikely to decline until TBT is banned on all vessels (Minchin et al., 1995). Even then, there is the possibility of continued contamination as TBT is persistent in sediments (Bryan & Gibbs, 1991; Hawkins et al., 1994).

Relevant additional considerations

Sufficiency of data: There is a considerable body of information on *N.lapillus* populations as well as changes in population numbers following the discovery of a link between TBT contamination and imposex. These studies continue, and have shown recovery of the populations in some areas as well as no improvement in other areas.

Changes in relation to natural variability: The significant decline in *N.lapillus* populations reported in the 1980s and 1990s have been linked to TBT contamination rather than the result of natural fluctuations in population numbers. A reduction in recruitment caused by a lowered reproductive capacity, therefore appears to be responsible for the decline in *N.lapillus* numbers.

Expert judgement: A link between decline in *N.lapillus* populations, imposex, and TBT has been demonstrated clearly, both in the field and in the laboratory. There have also been documented declines in populations following oil spills and toxic blooms. Consideration of the case on the basis of expert judgement has therefore not been necessary.

3. Current status of the species

Distribution of the feature in the OSPAR maritime area

N.lapillus is distributed from Iceland in the north, to Portugal in the south, and includes the Irish Sea and North Sea coasts (i.e. all OSPAR Regions).

Where suitable substrate is present, it can be ubiquitous. The species is wide spread on the rocky shores of Iceland, the United Kingdom, Ireland, France, and Norway. It also occurs on some sites of the north-west coast of Denmark, around the island of Helgoland (Germany) and in the south-west of the Netherlands.

N.lapillus has disappeared from Belgium since 1981. The species is reported to be absent in Sweden.

Population (current/trends/future prospects)

The size of populations is hard to assess. The distribution along the coastline may be very patchy, and juveniles and adults may hide in different places (cracks). Furthermore, individuals migrate and different age groups may be found at different tidal heights at different seasons. These factors, among others, make it very difficult to make repeatable estimations of population densities or population size. This accounts for all hard substrate species, in general. However, the presence of populations may be well noticed.

Declines of *N.lapillus* populations have been noted from the 1970s. The species has completely disappeared from Belgium in 1981 (Kerckhof, 1988), and declines were reported for other countries (e.g. Bryan *et al.*, 1986; Herbert, 1988).

Following bans on the use of TBT in the 1980s in several countries (see Section 5), the species has re-colonised several locations, on several other locations previously declining populations have (partially) recovered since the 1990s (e.g. Colson and Hughes, 2004; Huet *et al.*, 2004; Birchenough *et al.*, 2002; Moore *et al.*, 2000; Bray and Herbert, 1998; Evans *et al.*, 1994, 1995, 1996). Recolonisation may be slow as a result of the reproductive cycle of *N.lapillus*. Because juveniles emerge from egg capsules laid on the shore, their dispersal capability is poor. The species lacks a planktonic larval stage that may facilitate the re-colonisation of suitable substrates where TBT levels have declined. See annex 1 for trends per country.

Condition (current/trends/future prospects)

Little information on population quality aspects, such as age and size structure, is available. The most widely measured aspect is the level of imposex in female *N.lapillus*. Imposex has been observed in *N.lapillus* from the coastal areas of all countries bordering the North Sea, the Atlantic coast of Spain and Portugal, as well as in the more remote northern shores around Iceland (see Section 4). Following the bans on TBT (see Section 5) imposex levels in many locations have been declining since the 1990s (e.g. Huet *et al.*, 2004; Birchenough *et al.*, 2002; Miller *et al.*, 1999; Evans *et al.*, 1996; Harding *et al.*, 1997). The ban on TBT, however, appears to have taken much more rapid effect in the UK than in other countries such as France, Portugal and Spain (Ruiz *et al.*, 2005). In Portugal levels of imposex remained high and increasing until 2000, since then occurrence of imposex has also started to decrease there (Santos *et al.*, 2002).

However, areas frequented by large vessels and sites in the proximity of large harbours are still 'hot spots' of TBT contamination (e.g. Minchin *et al.*, 1996; Morgan *et al.*, 1998; Jorundsdottir *et al.*, 2005; Galante-Oliviera *et al.*, 2006). Most of these 'hot spots' are associated with commercial ports that require consistent maintenance dredging to ensure vessels can navigate freely (Galante-Oliviera *et al.*, 2006). This can lead to the secondary impact of TBT being released from sediments during dredging and spoil-disposal operations (Santos *et al.*, 2004).

Information on the impact of TBT on *N.lapillus* at the individual organism level (imposex) will become available from the monitoring activities under the OSPAR's Coordinated Environmental Monitoring Programme (CEMP) (see OSPAR publication 2009/390 (in press)).

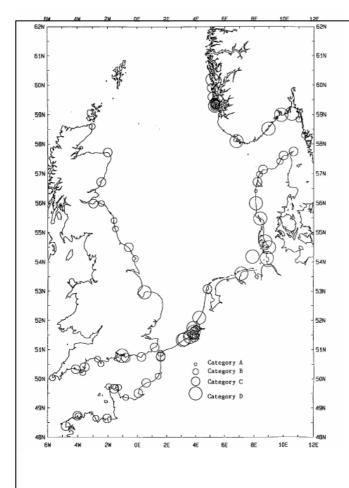


FIGURE 1: Categorisation of sites along the North Sea coast (from Harding *et al.*, 1999).

Category A: *N.lapillus* populations showing little, if any, effect of exposure to TBT beyond that associated with areas distant from sources of TBT (VDSI <2).

Category B: *N.lapillus* populations showing some effects of exposure to TBT, but not to the extent that any significant effect on production of egg capsules would be expected (2< VDSI <4).

Category C: *N.lapillus* populations showing more marked effects of exposure to TBT, to the extent that reduction in the production of egg capsules would be expected (4< VDSI <5).

Category D: *N.lapillus* populations showing severe effects of exposure to TBT, to the extent that production of egg capsules would be prevented (VDSI <5).

Limitations in knowledge

Sources of information on the distribution of *N.lapillus* have been provided by some Contracting Parties in the form of publications. These publications provide overviews of the results of monitoring programmes or surveys for entire coastlines (national level) or regions, and may concern the distribution of the species and/or the health status as reflected by imposex (VDSI level). Since *N.lapillus* is absent or rare in some countries, information is lacking.

Considering the proposed conservation objective, information needed for a full assessment includes:

- a description of the current and past distribution of *N.lapillus*. Information on the absence or presence of *N.lapillus* at suitable substrate (rocky shores and artificial hard substrates) would be sufficient for the assessment of the distribution on a wider spatial scale.
- information on the imposex level at locations around point sources of TBT.

4. Evaluation of threats and impacts

Threat and link to human activities

Relevant human activities: Shipping and navigation; tourism and recreational activities;

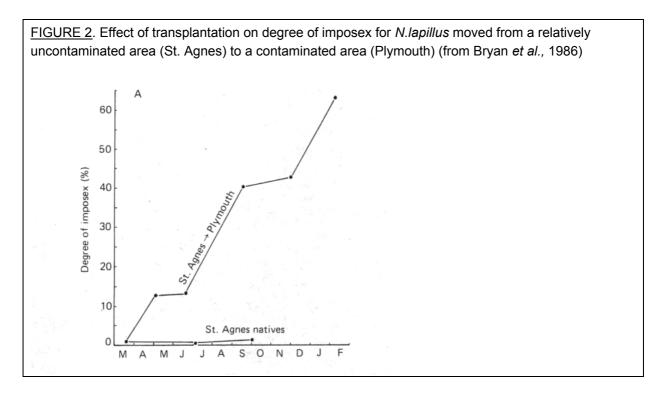
Category of effect of human activity: Chemical – synthetic compound contamination.

A direct link has been made between the decline in *N.lapillus* populations and the concentration of TBT in surrounding waters. There is evidence from field observations and laboratory studies that organotins originating from the TBT compounds used in antifouling paints cause imposex in *N.lapillus*, even at very low concentrations (e.g. Bryan *et al.*, 1986; Harding *et al.*, 1988; Gibbs *et al.*, 1991; Ruiz *et al.*, 1998; Santos *et al.*, 2005).

Further evidence for the relationship between imposex and TBT comes from transplantation experiments where *N.lapillus* were moved to areas where there was a high seawater concentration of tin (Smith *et al.*, 2006; Quintela *et al.*, 2000; Bryan *et al.*, 1986). This resulted in a gradual increase in the degree of imposex (Figure 2).

There is an increasing concern over the release of TBT from sediments through dredging or similar disturbance (Brack, 2002; Bray & Langston, 2006). Maritime ports require consistent maintenance dredging to ensure increasingly larger vessels to navigate freely. The impact of dumped dredged materials and resuspension may impact previously unaffected areas. In contrast to TBT in water, which was disappearing with a half-life of some three years after legislation, TBT in sediments is far more persistent and it is suggested that sediments may cause a problem for perhaps 20 to 30 years after a total ban (Macguire, 2000; Dowson *et al.*, 1996; Langston *et al.*, 1994). Dumping sites of dredged material containing TBT have been shown to have impact on imposex in gastropod populations (Santos *et al.*, 2004).

Oil pollution on rocky shores, and subsequent clean up operations are another potential threat to *N.lapilllus* populations (e.g. IPIECA, 1995). Declines have been observed following contamination of rocky shore with varying times for recovery depending on factors such as the severity of the spill, type of contamination, exposure of the shore, weather conditions and status before the incident (e.g. Bryan, 1968; Baker, 1976).



5. Existing management measures

Several measures have been taken to reduce the input of TBT into the marine environment. The use of TBT-based paints on vessels under 25 m was first banned by France in 1982 and was followed by other European countries between 1987 and 1991 with a similar ban throughout the North Sea

In 2001 the International Maritime Organisation (IMO) adopted 'The International Convention on the Control of Harmful Anti-fouling Systems on Ships' (AFS Convention) to work on a global legal instrument to ban TBT. The AFS Convention entered into force in 2008 and bans both the application and the presence on ships' hulls of TBT-based antifoulings.

In the meantime the European Union has, through Regulation (EC) No. 782/2003, banned the application of TBT-based paints on EU-flagged vessels and as of 1 January 2008 it is an offence for any ship visiting an EU port to have TBT present on its hull.

Another measure concerns the maximum concentrations of TBT for the dumping of dredged materials. According to the Overview of Contracting Parties' National Action Levels for Dredged Materials (OSPAR publication number 2004/211, updated in 2008 (publication number: 2008/363)), each country applies their own action levels concerning concentrations of TBT. OSPAR's Guidelines for the Management of Dredged Material (OSPAR agreement: 2009/4), which introduced the requirement of analyses of organic contaminants, came into force only in June 1998 (the Guidelines were revised in 2009).

IMO Convention	Effective Date	EC Regulation	Effective Date
Ships shall not apply or re-apply	1 January 2003	Ships shall not apply or reapply	1 July 2003
organotin compounds which act as		Organotin compounds which act as	
biocides in anti-fouling systems		biocides in anti-fouling systems	
Ships shall not bear such	1 January 2008	Ships whose anti-fouling system	1 July 2003
compounds on their hulls or		has been applied, changed or	
external surfaces; or shall bear a		replaced after the relevant date	
coating that forms a barrier to such		shall not bear such compounds on	
compounds leaching from the		their hulls or external surfaces; or	
underlying non-compliant anti-		shall bear a coating that forms a	
fouling systems		barrier to such compounds	
		leaching from the underlying non-	
		compliant anti-fouling systems	
		Ships shall not bear such	1 January
		compounds on their hulls or	2008
		external surfaces; or shall bear a	
		coating that forms a barrier to such	
		compounds leaching from the	
		underlying non-compliant anti-	
		fouling systems	

6. Conclusions on overall status

Since the introduction of a ban on use of TBT on small craft and subsequent bans on the use of TBT on larger vessels, some populations have recovered or have re-colonised places where they had disappeared (e.g. Evans *et al.*, 1994, 1995; Moore *et al.*, 2000; Huet *et al.*, 2004; Jorundsdottir *et al.*, 2005). A recent TBT Imposex assessment (OSPAR publication number: 2009/390) shows a significant downward trend in the occurrence of imposex in 24 populations versus for which there is more than 4 years of data and a significant upward trend in only 4 populations In 84 of the remaining 106 time series, the estimated trend was downwards, albeit non-significant.

Nevertheless even after the introduction of restrictions on the use of TBT, biological effects are still evident many areas, although often at lower levels than some years previously (Harding *et al.*, 1998. Areas frequented by large vessels and sites in the proximity of large harbours are still 'hot spots' of TBT contamination.

Following the worldwide ban on TBT-containing paints on all ships, levels of TBT are expected to decline further. Therefore, an improvement in the condition of populations of *N.lapillus* is expected over the next 10 years. TBT releases from sediment may, however, influence *N.lapillus* populations close to 'hot spots' or dumping sites for dredged material for a much longer period.

7. What action should be taken at an OSPAR level?

Action/measures that OSPAR could take, subject to OSPAR agreement

Actions by OSPAR so far:

Monitoring of TBT–specific biological effects in *N.lapillus* and other gastropods has been a mandatory element of the CEMP since 2003 for the purpose of tracking the implementation of the International Antifouling Substances Convention and related European provisions.

Marine Protected Areas (MPAs): The following MPAs reported to the MPA database include *N.lapillus:* Sept Îles (Fr); Gullmarn fjord (Sw); Las Islas Atlánticas (Sp)

Proposed further actions/measures:

Since an international ban is already in place, OSPAR should mainly focus on ensuring the enforcement of the worldwide ban on TBT.

Monitoring: OSPAR should urge Contracting Parties to improve data collection and evaluation of the CEMP monitoring programme.

Research: OSPAR should emphasise to relevant scientific funding bodies the need for further research into the release of TBT from dredged materials and the impact on *N.lapillus*. Following this research it may be necessary to adjust the national action levels for dredged materials for TBT.

Brief Summary of the proposed monitoring system

The proposed monitoring system includes both the spatial distribution and the condition of *N.lapillus*.

The proposed monitoring system builds further on the monitoring strategy described in the Joint Assessment and Monitoring Programme (JAMP) Guidelines for contaminant-specific biological effects monitoring (OSPAR agreement 2008-9). Contracting Parties where *N.lapillus* is rare or absent can instead monitor other gastropod species that are also affected by TBT. However, these species are less sensitive than *N.lapillus*.

The JAMP guideline focuses on the VDSI as the parameter to be monitored. However, at some locations *N.lapillus* may not be present at all as a result of exposure to TBT. Therefore, the presence and absence should be included.

The selection of locations is based on the following criteria:

- Populations of *N.lapillus* are currently absent, where they have been formerly present;
- *N.lapillus* populations are (potentially) exposed to increased concentrations of TBT.

Since the threats and decline resultfrom TBT in antifouling, the locations to be monitored are around major marinas, shipyards, offshore installations and harbours, as is stated in the Technical Annex 3 (TBT-specific biological effect monitoring) of the JAMP Guidelines (OSPAR Ref. No: 2008-9).

Annex 3 of the JAMP Guidelines provides instructions for the selection of monitoring locations around point sources and in the surrounding areas, the measurement of temporal trends, field sampling, and the determination of imposex.

Annex 1. Overview of data and information provided by Contracting Parties

Contracting Party	Feature occurs in Contracting Party's Maritime Area	Contribution made to the assessment (e.g. data/information provided)	National reports References or weblinks
Belgium	No	providedy	Kerckhof, F. (1988). Over het verdwijnen van de Purperslak Nucella lapillus (Linnaeus, 1758) langs de Belgische kust. De Strandvlo 8(2): 82-85.
distribut		Information on distribution and monitoring	Ris, M.A. (1930). Den danske strand. Naturforhold, dyre- og planteliv ved vore ky-ster. C.A.Reitzels forlag, København, Danmark, 139pp (in Danish)
		programme	Strand, J. 2003. Coupling marine monitoring and risk assessment by integrating exposure, bioaccumulation and effect studies. A case study using the contamination of organotin compounds in the Danish marine environment. PhD thesis. Roskilde University. 92 pp + papers. Available at: http://hdl.handle.net/1800/571.
			Strand 1998, Strand 2004, Strand unpubl.
			Faroe islands: Følsvik, N., E.M. Brevik, J.A. Berge & M. Dam (1998). Organotin and imposex in the Littoral Zone in the Faroe Islands, Fróðskaparrit. 46: 67-80.
European Commission			See report of the EU SCTEE with a chapter on imposex in marine snails including <i>N.lapillus</i> : Vos, JG, Dybing, E Greim, HA, Ladevoged, O, Lambré, C, Tarazona, JV, Brandt, I, Vethaak, AD (1999). CSTEE Opinion on Human and Wildlife Health Effects of Endocrine Disrupting Chemicals, with Emphasis on Wildlife and on Ecotoxicity Test Methods. Report of the Working Group on Endocrine Disrupters of the Scientific Committee on Toxicity, Ecotoxicity and the Environment (CSTEE) of DG XXIV, Consumers Policy and Consumer Health Protection, 96 pp. http://europe.eu.int/comm/dg24/health/sc/sct/out37_en.html
France	Yes	Information on distribution and monitoring programme.	Huet M. & Paulet Y-M. 2006. Estimation de la pollution par le tributyletain en 2006 a l'aide de l'imposex. Réseau National d'Observation. Huet M., Y. M. Paulet & J. Clavier. 2004. Imposex in <i>Nucella</i>
		Time series for Imposex.	lapillus: a ten year survey in NW Brittany. Mar Ecol Prog Ser. Vol. 270: 153–161.

Germany	Yes	Information on distribution	
Iceland	Yes	Information on distribution and monitoring	Ingolfsson, A. 1996. The distribution of intertidal macrofauna on the coasts of Iceland in relation to temperature. Sarsia 81: 29-44)
		programme.	Ingolfsson, A. 2006. The intertidal seashore of Iceland and its animal communities. Zoology of Iceland 1(7): 1-85.
			Jörundsdóttir K., Svavarsson J. & Leung K.M.Y. 2005. Imposex levels in the dogwhelk <i>Nucella lapillus</i> (L.)— continuing improvement at high latitudes. Marine Pollution Bulletin 51 (2005) 744–749,
Ireland	Yes	Information on distribution and monitoring programme.	
Netherlands	Yes	Information on distribution and monitoring programme.	Gmelig Meyling A.W., J. Willemsen & R.H. de Bruyne. 2006. Verspreiding en trend Purperslak <i>Nucella lapillus</i> . Stichting Anemoon. Available at: http://www.anemoon.org/anemoon/downloads/rapporten/PIM P_2006_10_20.pdf
			Gmelig Meyling A.W., Borren,H. & J.Willemsen. 2007. Purperslak <i>Nucella lapillus</i> Inventarisatie en Monitoringproject Jaarverslag 2007. Stichting Anemoon. Available at: http://www.anemoon.org/anemoon/downloads/rapporten/PIM P_2007.pdf
			Cor A. Schipper, Mathijs G.D. Smit, Nicholas H.B.M. Kaag, A. Dick Vethaak, (2008), A weight-of-evidence approach to assessing the ecological impact of organotin pollution in Dutch marine and brackish waters; combining risk prognosis and field monitoring using common periwinkles (<i>Littorina littorea</i>), Mar. Env. Res., 66, 231–239
Norway	Yes	Information on distribution and monitoring programme.	
		Time series for Imposex.	
Portugal	Yes		
Spain	Yes		
Sweden	No		
UK	Yes	Information on distribution and monitoring	

	programme.	
	Time series for	
	Imposex	

Nucella Lapillus was nominated for inclusion in the OSPAR List by Belgium and WWF in 2001. (Contact Persons: Belgium: Jan Haelters & Francis Kerckhof, Management Unit of the North Sea Mathematical Models, 3e en 23e Linieregimentsplein, 8400 Oostende, Belgium. WWF: Sabine Christiansen, International WWF Centre for Marine Conservation, Hongkongstr.7., D-20457 Hamburg, Germany

Summaries of country-specific information provided

Belgium:

In Belgium *N.lapillus* has disappeared since 1981. Belgian expect that the species may re-colonise in the future from France on breakwaters along the coastline. *N.lapillus* is not included in a monitoring programme, information is received from volunteers.

Denmark:

Conservation of *N.lapillus* is not a national objective. *N.lapillus* only occurs on the west coast of Denmark from Hvide Sande to Skagen, i.e. the coastlines of the North Sea and the Skagerrak. Before 1930 *N.lapillus* was also found in the Kattegat. *N.lapillus* are found only on artificial rocks, i.e. harbour piers and breakwaters for protection of the coast.

N.lapillus is a widespread species in the tidal zone at the Faroe Islands. The species is very uncommon in Greenland.

Nucella is included in the Danish monitoring programme, but only as an imposex indicator. In 1991, 1998/1999, 2004 and 2007 imposex in N.lapillus was assessed at 2-6 stations, both sites close to harbours and sites regarded as reference sites. N.lapillus does not have a high priority in the national monitoring programme, since the monitoring of benthic communities is focused on the gastropods Buccinum undatum and Neptunea antiqua in open waters and Nassarius reticulatus and Littorina littorea in coastal areas.

France:

Conservation of *N.lapillus* is not a national objective. In France *N.lapillus* is widely distributed along the Atlantic coast. Recent trends show a decline in imposex and recolonisation of harbours (Brest). *N.lapillus* is included in the French monitoring programme only as an imposex indicator.

Germany:

Comprehensive data on the occurrence and abundance of *N.lapillus* in the German North Sea are limited. Rocky shores as suitable habitat for this species can only be found around the Island of Helgoland where *N.lapillus* is known to exist in small numbers. Hence, in Germany the species is included in the "Red List of Threatened Species" and categorized as "very rare"; with no signs of short-term positive changes in its abundance. *N.lapillus* is considered to be "highly threatened" (Category 2).

Iceland:

Conservation of *N.lapillus* is not a national objective. *N.lapillus* is common on the west coast of Iceland. The species had previously disappeared from one site, but has recently reappeared. There is a substantial recovery of the species near small harbours, recovery is slower but still continuing near larger harbours.

N.lapillus is included in the monitoring programme. Presence, length, distribution and imposex are monitored.

Ireland:

Conservation of *N.lapillus* is not a national objective. *N.lapillus* occurs everywhere along the coast on rocky substrates. *N.lapillus* is not yet included in the Irish monitoring programme, but imposex monitoring is under development for the Water Framework Directive.

Netherlands:

Conservation of *N.lapillus* is not a national objective. *N.lapillus* in the Netherlands only occurs on artificial substrates such as dykes and breakwaters. Its current distribution is limited to the south-west.

N.lapillus does not occur in the Wadden Sea region. In the North Sea coastal area, the number of localities with populations of *N.lapillus* has decreased between 1960 and 1997. At some of these localities, the habitat was drastically changed by silting up and the covering of basalt with asphalt/tar. In the Eastern Scheldt region, a marked decline in numbers occurred between 1970 and 1997. Numbers have recovered from 1998 onward. The populations once occurring in the mouth of the Western Scheldt have disappeared. This decline seems to have started in the 1960s.

N.lapillus is only monitored by volunteers. Time series data since the 1970s have been compiled. *Littorina littorea* is used as an imposex indicator because of its wider distribution.

Norway:

Conservation of *N.lapillus* is not a national objective. Distribution of the species has only been, semi-qualitatively assessed, along the southern coast.

Imposex in *N.lapillus* and TBT concentrations are included in the Annual Norwegian Coastal Monitoring Programme (NMCP)

Portugal:

No information

Spain:

No information

Sweden:

Sweden reports that *N.lapillus* is not present along the Swedish coast. For imposex monitoring *Nassarius reticulatus* is used instead.

UK:

Conservation of *N.lapillus* is not a national objective. *N.lapillus* occurs everywhere along the coast on rocky substrates. The condition of the population continues to improve. Re-colonisation of locations is low and depends on the distribution of existing populations.

Imposex has been monitored in the Marine National Monitoring Programme of the UK in 1991, 1997, 2003 and 2007. In the Shetlands, imposex has been monitored every 2 years for the past 20 years. The UK does not assess densities since it is very difficult to get repeatable estimations of hard substrate organisms for several reasons (patchy distribution, hide in cracks, accumulations of breeding and non-breeding individuals, juveniles hide in different places, seasonal differences).

Annex 2. Description of the recommended monitoring and assessment strategy

Rationale

The species is recovering after a strong decline in the 1970s. However, areas frequented by large vessels and sites in the proximity of large harbours are still 'hot spots' of TBT contamination and recovery here is very slow or absent. Most of these 'hot spots' are associated with commercial ports that require consistent maintenance dredging to ensure vessels can navigate freely. This can lead to the secondary impact of TBT being released from sediments during dredging and spoil-disposal operations.

Following the worldwide ban on TBT-containing paints on all ships, levels of TBT are expected to fall further and an improvement in the condition of populations of *N.lapillus* is expected during the next 10 years. Good monitoring programmes are already in place in most contracting parties under the CEMP. It is recommended to continue monitoring to follow the effects of the world-wide ban on TBT.

Use of existing monitoring programmes

The OSPAR Coordinated Environmental Monitoring Programme (CEMP) currently covers monitoring and assessment criteria of the EcoQO on imposex in *N.lapillus* and other selected gastropods.

Guidelines for the TBT-specific biological effects monitoring are described in the Technical Annex 3 of the JAMP Guidelines for Contaminant-specific Biological Effects Monitoring. (OSPAR Reference number 2008-9).

Synergies with monitoring of other species or habitats

N.lapillus is found on wave exposed to sheltered rocky shores. Its distribution is not shared by other intertidal rocky shore organisms or habitats on the OSPAR list of threatened and declining species and habitats.

The monitoring of the status of the population of *N.lapillus* may coincide with the monitoring of the EcoQO imposex in *N.lapillus*.

For the TBT-specific biological effect monitoring, alternative gastropod species may be used where *N.lapillus* does not occur. For rocky shores these species are *Nassarius reticulatus* and *Littorina littorea*, and for offshore areas *Buccinum undatum* and *Neptunea antiqua*.

The following table provides an overview of the monitoring Contracting Parties already have in place (table 4.9. from ICES, 2004b – WGBEC 2004).

Country/area	Programme type	Number of stations	Frequency	Number of years	Species	
BE	monitoring	3	annual	1	Littorina littorea	
DK	monitoring	15	every 2 yrs	5	Nassarius reticulatus	
DK	monitoring	14	every 2 yrs	6	Buccinum and/or Neptunea	
DK	monitoring	12	every 2 yrs	5	Littorina littorea	
FR	monitoring	108	annual	2	Nucella lapillus	
IS	monitoring		every 5 yrs	10	Nucella lapillus	
NL	monitoring	6	new	new	Littorina littorea	
NO	monitoring	15	every 4 yrs	12	Nucella lapillus Nassarius reticulatus Nucella lapillus Buccinum undatum Nucella lapillus	
SE	monitoring	5	new	new		
UK	monitoring (hotspot)	8	every 5 yrs	15		
UK	monitoring (offshore)	21	new	new		
UK	monitoring (spatial)	82	every 5 yrs	10		

Assessment criteria

The only agreed conservation objective concerns the level of imposex in gastropods (VDSI<2). *N.lapillus* is the preferred species for the monitoring of imposex since it is most sensitive. However, if the species is rare or absent other species may be used.

The following criteria for imposex / intersex related to the Environmental Assessment Criteria and effects on reproductive capacity for gastropod species have been established (OSPAR Reference Number 2004-15):

The Ecological Quality Objective relates to a VDSI for *N.lapillus* <2.

No conservation objectives on the distribution of populations have been established.

Assessment	Nucella	Nassarius	Buccinum~	Neptunea#	Littorina
class	VDSI	VDSI	PCI	VDSI	ISI
А	< 0.3	< 0.3 ¹	< 0.3 ¹	< 0.3	< 0.3 ²
В	0.3 - <2.0			0.3 - <2.0	
С	2.0 < 4.0	0.3 < 2.0	0.3 < 2.0	$2.0 < 4.0^3$	
D	4.0 - 5.0	2.0 - 3.5	2.0 - 3.5	4.0 ³	0.3 - < 0.5
Е	>5.0	> 3.5 ⁴	> 3.5 ⁴		0.5 - 1.2
F	-				> 1.2
		Stroben <i>et al.</i> , 1995	Stroben et al., 1995, and field evidence that Buccinum has similar sensitivity as Nassarius, ~ No correlation established	# field evidence that Neptunea has similar sensitivity as Nucella, ^ highest value possible ~ No correlation established	Oehlmann, 2002 ASMO 02/4/8

¹ This species cannot be used to distinguish between class A and class B. The assessment class is therefore by definition B.

Techniques/approaches:

Baseline

Baseline requirements would include the implementation of the CEMP monitoring, i.e. TBT-specific biological effects monitoring, around point sources and regional surveys.

Enhanced

Since exposure to TBT may have lead to the complete disappearance of *N.lapillus* population<u>s</u>, the monitoring of the spatial presence (or absence) of *N.lapillus* will give insight in the locations that are still under threat of high TBT concentrations, or where re-colonisation has not (yet) occurred because of the absence of populations in the vicinity of these locations.

Additional requirements would be the chemical analysis of TBT in tissue of *N.lapillus*, and if desired in other environmental compartments (water, suspended matter, sediment). This may help to identify the main sources of threat.

Selection of monitoring locations

Guidance on the selection of monitoring locations is provided by the CEMP. The selection includes both impacted and unimpacted locations.

² This species cannot be used to distinguish between classes A, B and C. The assessment class is therefore by definition C.

³ This species cannot be used to distinguish between class C and higher classes. If a VDSI of 4.0 is reached, additional observations are required to determine the assessment class e.g. by using another species. If a VDSI of 4.0 is observed, the assessment class is therefore by definition F.

⁴ These species cannot be used to distinguish between class E and class F. Therefore, additional observations are required to determine the assessment class e.g. by using another species If the VDSI (*Nassarius*) or the PCI (*Buccinum*) is >3.5, the assessment class is therefore by definition F.

Timing and Frequency of monitoring.

Since the monitoring should be able to show improvement of the condition of *N.lapillus*, monitoring activities should be able to detect temporal changes. It is therefore recommended to sample locations for imposex at least every 2 years. It would be better to select a limited number of locations that are monitored frequently, than to select a high number of locations that can only be monitored at a low frequency (e.g. once in 5 years).

Spatial distribution (at a regional level) could be monitored less frequently, since the recolonisation rate of *N.lapillus* is rather slow.

Data collection and reporting

The OSPAR JAMP guidelines for Contaminant-specific biological effects monitoring (OSPAR agreement 2008-9) provides the following reporting requirements:

Data reporting should be in accordance with the requirements for national comments and with the latest ICES reporting formats. The following data are required:

Contaminants

- TBT, DBT and MBT concentrations in tissues (mg/kg);
- TPhT, DPhT, MPhT concentrations in tissues (mg/kg) when relevant;
- wet weight or dry weight basis;

Biological effects measurements

Imposex

- proportion of females displaying imposex;
- · vas deferens sequence index;
- · relative penis size
- proportion of sterile females in stages 5 and 6.

Supporting parameters

- site code:
- · taxonomic identification;
- · date of sample collection;
- number of individuals in sample
- presence/absence of juveniles and/or egg capsules;
- population size & frequency distribution, if considered useful supplementary information.

Quality assurance

International Laboratory Performance Studies of imposex (and intersex for other species than *N. lapillus*) measurements are available through QUASIMEME and provide a formal framework for external quality assurance.

Annex 3. References

OSPAR Publications:

OSPAR 2004/211. Overview of Contracting Parties' National Action Levels for Dredged Material.

OSPAR 2004-15. Provisional JAMP Assessment Criteria for TBT – Specific Biological Effects. OSPAR Commission 2004.

OSPAR 2005. North Sea Pilot Project on Ecological Quality Objectives. Background Document on the Ecological Quality Objective on imposex in dog whelks *Nucella lapillus*. OSPAR Commission 2005.

OSPAR 2006-239. Report on North Sea Pilot Project on Ecological Quality Objectives. OSPAR Commission 2006.

OSPAR 2008-358. Case Reports for the OSPAR List of Threatened and/or Declining Species and Habitats in the OSPAR Maritime Area. OSPAR Commission 2006.

OSPAR 2008-9. JAMP Guidelines for Contaminant-specific Biological Effects Monitoring. OSPAR Commission 2008.

OSPAR 2008/363. Update of the Overview of Contracting Parties' National Action Levels for Dredged Material

OSPAR 2009-1. Revised OSPAR Coordinated Environmental Monitoring Programme (CEMP).

OSPAR 2009/390. CEMP assessment report: 2008/2009. Assessment of trends and concentrations of selected hazardous substances in sediments and biota

Other References:

Baker, J.M., (1976). Ecological changes in Milford Haven during its history as an oil port. In *Proceedings of an Institute of Petroleum / Field Studies Council meeting, Aviemore, Scotland, 21-23 April 1975. Marine Ecology and Oil Pollution* (ed. J.A. Baker), pp. 55-66. Barking: Applied Science Publishers Ltd.

Birchenough, A.C., Evans, S.M., Moss, C. & Welch, R. (2002) Re-colonisation and recovery of populations of dogwhelks Nucella lapillus (L.) on shores formerly subject to severe TBT contamination. *Marine Pollution Bulletin.* 44 (7); 652-659.

Blaber, S.J.M. (1970). The occurrence of a penis-like outgrowth behind the right tentacle in spent females of *Nucella lapillus* (L.) *Proc.Malacol.Soc.Lond*. 39:231-233.

Brack, K. (2002) Organotin compounds in sediments from the Gota Alv estuary. *Water, Air and Soil Pollution*. 135; 131-140.

Bray, S. and Herbert, R.J.H. (1998) A reassessment of populations of the dog-whelk (*Nucella lapillus*) on the Isle of Wight following legislation restricting the use of TBT antifouling paints. *Proceedings of the Isle of Wight Natural History and Archaeological Society.* 14; 23-40.

Bray, S. and Langston,B. (2006). Tributyltin pollution on a global scale. An overview of relevant and recent research: impacts and issues. WWF, Godalming, UK.

Bryan, G.W., (1968). The effect of oil-spill removers ('detergents') on the gastropod *Nucella lapillus* on a rocky shore and in the laboratory. *J.Mar.Biol.Ass.UK.*, 49:1067-1092.

Bryan, G.W., Gibbs, P.E., Hummerstone, L.G. & Burt, G.R. (1986).. The decline of the gastropod *Nucella lapillus* around south-west England: evidence for the effect of tributyltin from antifouling paints. *J.Mar.Biol.Ass.UK*. 66(3); 611-640.

Bryan, G.W. & Gibbs, P.E., (1991). Impact of low concentrations of tributyltin (TBT) on marine organisms: a review. In: *Metal ecotoxicology: concepts and applications*, (ed. M.C. Newman & A.W. McIntosh), pp. 323-361. Boston: Lewis Publishers Inc.

Colson, I. and Hughes, R.N. (2004) Rapid recovery of genetic diversity of dogwhelk (*Nucella lapillus* L.) populations after local extinction and recolonisation contradicts predictions from life history characteristics. *Molecular Ecology*. 13; 2223-2233.

Cross, T.F. & Southgate, T., (1980). Mortalities of fauna of rocky substrates in south-west Ireland associated with the occurrence of *Gyrodinium aureolum* blooms during autumn 1979. *J.Mar.Biol.Ass.* 60:1071-1073.

Dowson, P.H., Bubb, J.M. and Lester, J.N. (1996) Persistence and degradation pathways of tributyltin in freshwater and estuarine sediments. Estuarine Coastal and Shelf Science. 42 (5); 551-562.

Evans, S.M., Evans, P.M. & Leksono, T., (1996). Widespread recovery of dogwhelks, *Nucella lapillus* (L.), from tributyltin contamination in the North Sea and Clyde Sea *Mar.Poll.Bull.*, 32, 263-369.

Evans, S.M., Hawkins, S.T., Porter, J. & Samosir, A.M. (1994). Recovery of dogwhelk populations on the Isle of Cumbrae, Scotland, following legislation limiting the use of TBT as an antifoulant. *Mar.Poll.Bull.* 28(1): 15-17.

Evans, S.M., Kerrigan, E. & Palmer, N. (2000). Causes of imposex in the dogwhelk *Nucella lapillus* (L.) and its use as a biological indicator of tributyltin contamination. *Mar.Poll.Bull.* 40:212-219.

Evans, S.M., Leksono, T. & McKinnell, P.D. (1995). Tributyltin pollution: a diminishing problem following legislation limiting the use of TBT-based anti-fouling paints. *Mar.Poll.Bull.* 30(1): 14-21.

Galante-Oliveira, S., Langston, W.J., Burt, G.R., Pereira, M.E. and Barroso, C.M. (2006). Imposex and organotin body burden in the dog-whelk (*Nucella lapillus* L.) along the Portuguese coast. *Applied Organometallic Chemistry* 20; 1-4

Gibbs, P.E. & Bryan, G.W. (1987). The use of the dogwhelk *Nucella lapillus*, as an indicator of tributyltin (TBT) contamination. *J.Mar.Biol.Ass. UK*. 67:507-523.

Gibbs, P.E. & Bryan, G.W. (1996). Reproductive failure in the gastropod *Nucella lapillus* associated with imposex caused by tributyltin pollution: a review. In Organotin, Environmental Fate and Effects (Eds) Champ, M.A. & Seligman, P.F. P259-280. Chapman & Hall, London.

Gibbs, P.E., Bryan, G.W. & Pascoe, P.L. (1991). TBT-induced imposex in the dogwhelk *Nucella lapillus*: geographical uniformity of the response and effects. *Mar.Env.Res.* 32:79-87.

Gibbs, P.E., Green, J.C. & Pascoe, P.C., (1999). A massive summer kill of the dog-whelk, *Nucella lapillus*, on the north Cornwall coast in 1995: freak or forerunner? *J.Mar.Biol.Ass.UK.*, 79, 103-109.

Harding, M.J.C., Davies, I.M., Bailey, S.K. and Rodger, G.K. (1999) Survey of Imposex in Dogwhelks (*Nucella lapillus*) from North Sea Coasts. *Applied Organometallic Chemistry* 13; 521–538

Harding, M.J.C., Davies, I.M., Minchin, I.M & Grewar, G. (1988). Effects of TBT in western coastal waters. PECD CW0691. Fisheries Research Services Report. No.5/98. Scottish Office Agriculture, Environment and Fisheries Department 39pp.

Harding, M.J.C., Rodger, G.K., Davies, I.M. and Moore, J.J. (1997) Partial recovery of the dogwhelk (*Nucella lapillus*) in Sullom Voe, Shetland, from tributyltin contamination. *Marine Environmental Research*. 44 (3); 285-304.

Hawkins, S.J., Proud, S.V., Spence, S.K. & Southward, A.J., (1994). From the individual to the community and beyond: water quality, stress indicators and key species in coastal systems. In *Water quality and stress indicators in marine and freshwater ecosystems: linking levels of organisation (individuals, populations, communities)* (ed. D.W. Sutcliffe), 35-62. Ambleside, UK: Freshwater Biological Association.

Herbert, R.J.H. (1988) A survey of the dogwhelk Nucella lapillus (L.) around the coast of the Isle of Wight. *Proceedings of the Isle of Wight Archaeological and Natural History Society*. 8 (3); 15-21.

Huet M., Y. M. Paulet & J. Clavier. 2004. Imposex in *Nucella lapillus*: a ten year survey in NW Brittany. *Mar Ecol Prog Ser*. 270: 153–161.

IPIECA (1995). Biological Impacts of oil pollution: rocky shores. IPIECA Report Series, Vol.7. International Petroleum Industry Environmental Conservation Association, London.

Jörundsdóttir K., Svavarsson J. & Leung K.M.Y. 2005. Imposex levels in the dogwhelk *Nucella lapillus* (L.)—continuing improvement at high latitudes. *Marine Pollution Bulletin* 51 (2005) 744–749, see MON 07/2/Info.1

Kerckhof, F. (1988). Over het verdwijnen van de purperslak Nucella lapillus (L.1758). langs onze kust. De Strandvlo, 8(2): 82-85.

Langston, W.J., Bryan, G.W., Burt, G.R. and Pope, N.D. (1994) *Effects of Sediment Metals on Estuarine Benthic Organisms*. National Rivers Authority R and D Note 203. NRA, Almondsbury, Bristol.

Maguire, R.J. (2000) Review of the persistence, bioaccumulation and toxicity of tributyltin in aquatic environments in relation to Canada's toxic substances management policy. *Water Quality Research Journal of Canada*. 35 (4); 633-679.

Matthiessen, P. and Gibbs P.E. (1998) Critical appraisal of the evidence for tributyltinmediated endocrine disruption in molluscs. *Environmental Toxicology and Chemistry*. 17; 37–43.

Miller, K.L., Fernandes, T.F. and Read, P.A. (1999) The recovery of populations of dogwhelks suffering from imposex in the Firth of Forth 1987-1997/98. *Environmental Pollution*. 106; 183-192.

Minchin, D., Oehlmann, J., Duggan, C.B., Stroben, E., & Keatinge, M. (1995). Marine TBT antifouling contamination in Ireland, following legislation in 1987. *Mar.Poll.Bull.* 30(10): 633-639.

Moore, J.J., James, B., Minchin, A. & Davies, I.M., (2000). Surveys of dog whelks *Nucella lapillus* in the vicinity of Sullom Voe, Shetland, August 1999. *Report to the Shetland Oil Terminal Environmental Advisory Group (SOTEAG)*, prepared by CORDAH Ltd and the Fisheries Research Services.

Morgan, E., Murphy, J. & Lyons, R. (1998). Imposex in *Nucella lapillus* from TBT contamination in south and south-west Wales: a continuing problem around ports. *Mar.Poll.Bull.* 36(10): 840-843.

OSPAR (2000). Quality Status Report 2000. Region III. Celtic Seas. OSPAR Commission, London. 116pp.

Quintela, M., Barreiro, R. and Ruiz, J.M. (2000) The use of Nucella lapillus (L.) transplanted in cages to monitor tributyltin (TBT) pollution. *The Science of The Total Environment*. 247 (2-3); 227-237.

Robertson, A., (1991). Effects of a toxic bloom of *Chrysochromulina polylepis*, on the Swedish west coast. *J.Mar.Biol.Ass.UK*, 71, 569-578.

Ruiz, J.M., Barreiro, R. and González, J.J. (2005) Biomonitoring organotin pollution with gastropods and mussels. *Marine Ecology Progress*. 287; 169-176.

Ruiz, J.M., Quintela, M. & Barreiro, R. (1998). Ubiquitous imposex and organotin bioaccumulation in gastropods *Nucella lapillus* (L.) from Galicia (NW Spain): a possible effect of nearshore shipping. *Mar.Ecol.Prog.Ser.* 164:237-244.

Santos, M.M., Castro, L.F.C., Vieira, M.N., Micael, J.; Morabito, R., Massanisso, P. and Reis-Henriques, M.A (2005) New insights into the mechanism of imposex induction in the dogwhelk *Nucella lapillus*. *Comparative Biochemistry and Physiology Part C Toxicology and Pharmacology*. 141(1); 101-109

Santos, M.M., Ten Hallers-Tjabbes, C.C., Santos, A.M. and Viera, N. (2002) Imposex in *Nucella lapillus*, a bioindicator for TBT contamination: resurvey along the Portuguese coast to monitor the effectiveness of EU regulation. *Journal of Sea Research*. 48; 217-223.

Santos, M.M., Viera, N. Reis-Henriques, M.A., Santos, A.M., Gomez-Ariza, J.L., Giraldaez, I. and ten Hallers-Tjabbes, C.C (2004) Imposex and butyltin contaminants off the Oporto Coast (NW Portugal): a possible effect of the discharge of dredged material. *Environment International*. 30; 793-798.

Skarphédinsdóttir, H., Ólafsdóttir, K, Svavarsson, J, & Jóhannesson, T. (1996). Seasonal fluctuations of Tributyltin (TBT) and Dibutyltin (DBT) in the Dogwhelk, *Nucella lapillus* (L.) and the Blue mussel, *Mytilus edulis* L., in Icelandic waters. *Mar.Poll.Bull.* 32(4): 358-361.

Smith, A.J., Thain, J.E., and Barry, J.(2006) Exploring the use of caged Nucella lapillus to monitor changes to TBT hotspot areas: A trial in the River Tyne estuary (UK). *Marine-Environmental-Research*. 62(2); 149-163

Svavarsson, J & Skarphéoinsdóttir, H. (1995). Imposex in the dogwhelk *Nucella lapillus* (L.) in Icelandic waters. *Sarsia* 80:35-40.

Tyler-Walters, H. (2002). Dogwhelk, *Nucella lapillus*. *Marine Life Information Network: Biology and Sensitivity Key Information Sub-programme* [On-line]. Plymouth: Marine Biological Association of the United Kingdom. [cited 15th October 2002]. Available from: http://www.marlin.ac.uk

Underdal, B., Skulberg, O.M., Dahl, E & Aune, T., (1989). Disastrous bloom of *Chrysochromulina polylepis* (Pymnesiophycaea) in Norwegian Coastal Waters 1988 - mortality in marine biota. *Ambio*, 18,:265-270.



New Court 48 Carey Street London WC2A 2JQ United Kingdom t: +44 (0)20 7430 5200 f: +44 (0)20 7430 5225 e: secretariat@ospar.org www.ospar.org

OSPAR's vision is of a clean, healthy and biologically diverse North-East Atlantic used sustainably

ISBN 978-1-906840-48-8 Publication Number: 408/2009

© Commission OSPAR, 2009. La reproduction de tout ou partie de ce rapport dans une publication peut être autorisée par l'Editeur, sous réserve que l'origine de l'extrait soit clairement mentionnée.

[©] OSPAR Commission, 2009. Permission may be granted by the publishers for the report to be wholly or partly reproduced in publications provided that the source of the extract is clearly indicated.